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Lignin degradation in wood-feeding insects

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The aromatic polymer lignin protects plants from most forms of microbial attack. Despite the fact that a significant fraction of all lignocellulose degraded passes through arthropod guts, the fate of lignin in these systems is not known. Using tetramethylammonium hydroxide thermochemolysis, we show lignin degradation by two insect species, the Asian longhorned beetle (Anoplophora glabripennis) and the Pacific dampwood termite (Zoötermopsis angusticollis). In both the beetle and termite, significant levels of propyl side-chain oxidation (depolymerization) and demethylation of ring methoxyl groups is detected; for the termite, ring hydroxylation is also observed. In addition, culture-independent fungal gut community analysis of A. glabripennis identified a single species of fungus in the Fusarium solani/Nectria haematococca species complex. This is a soft-rot fungus that may be contributing to wood degradation. These results transform our understanding of lignin degradation by wood-feeding insects.

Asian longhorned beetle | Pacific dampwood termite | TMAH thermochemolysis | Anoplophora glabripennis | Zoötermopsis angusticollis

Lignin plays a central role in carbon cycling on Earth. Its heterogeneous structure imparts plants with structural rigidity and also serves to protect cellulose and hemicellulose from degradation (1). Most of what is known about lignin biodegradation is from pure culture studies with filamentous basidiomycete fungi, known as white-rot and brown-rot decay. Although both white-rot and brown-rot fungal degradation have been characterized, much more is known about the white-rot system (2, 3). White-rot fungi simultaneously degrade the three major components of the plant cell wall: lignin, cellulose, and hemicellulose. Analysis of white-rot–degraded wood shows that the reactions in lignin: (i) are oxidative, (ii) involve demethylation (or demethoxylation), (iii) include side-chain oxidation at Cα, and (iv) involve propyl side-chain cleavage between Cα and Cβ (Fig. 1) (4). In contrast to white-rot fungi, brown-rot fungi are able to circumvent the lignin barrier, removing the hemicellulose and cellulose with only minor modification to the lignin. Consequently, lignin remains a major component of the degraded plant cell wall (5). The remaining lignin is demethylated on aryl methoxy groups and contains a greater number of ring hydroxy groups (6).

Little is known about lignin degradation in complex ecosystems, such as insect guts, where a consortium of microbes may be involved in degradation rather than just a single species. Although cellulose degradation in insect guts is well documented (7, 8), the fate of lignin has not clearly been demonstrated (9, 10), and it is widely accepted that insect gut systems do not have the capacity to degrade lignin (10). Although the majority of previous reports suggest that many wood-feeding insects overcome the lignin barrier by feeding on predegraded wood (11) or through exosymbiotic relationships with wood-degrading fungi (12, 13), there are species of insects that feed on the inner wood of living, healthy trees. How these insects are able to circumvent the lignin barrier and gain access to the polymer carbohydrates has been a mystery. Herein, we show chemical changes in lignin upon passage of undegraded wood through two wood-feeding insect gut systems, one that naturally feeds on living healthy trees and a second that typically feeds on dead wood.

To determine the fate of lignin in these two systems, we analyzed the frass of both the Asian longhorned beetle (Anoplophora glabripennis) and the Pacific dampwood termite (Zoötermopsis angusticollis). A. glabripennis is a wood-feeding cerambycid beetle that develops in the inner wood of a wide range of hardwood trees (14), feeding on living intact wood. Z. angusticollis is a lower termite that feeds on coniferous trees, attacking dead timber. Lignin structure of the frass was analyzed by unlabeled and 13C-tetramethylammonium hydroxide (TMAH) thermochemolysis, depolymerizing lignin into its aromatic sub-units by breaking β-O-4 linkages and methylating all ring hydroxyls (15). The resulting product is then analyzed by GC/MS to determine how the lignin structure has been chemically modified. In undegraded wood, the predominant products from TMAH thermochemolysis are typically 3,4-dimethoxybenzaldehyde (G4) from guaiacyl (G) lignin (Fig. 1) and 3,4,5-trimethoxybenzaldehyde (S4) from syringyl (S) lignin. Hardwood tree species contain both guaiacyl and syringyl lignin, but coniferous trees only contain guaiacyl lignin.

Results

Studies with fungal lignin degradation show three predominant reactions: (i) propyl side-chain oxidation/cleavage (reaction 1, Fig. 1), (ii) ring hydroxylation (reaction 2, Fig. 1), and (iii) demethylation (reaction 3, Fig. 1) (15). All these processes can be detected by labeled or unlabeled TMAH analysis (Fig. 1). Side-chain oxidation is a predominant reaction with white-rot fungi resulting in Cα–Cβ cleavage/demethylation of lignin (16). Previous studies demonstrate that oxidative alteration of the lignin propyl side chain results in the enhanced production of 3,4-dimethoxybenzoic acid (G6) and 3,4,5-trimethoxybenzoic acid (S6) (for hardwoods) (Fig. 2), and increases in G6/G4 and S6/S4 ratios indicate a higher degree of side-chain oxidation (16). Frass (feces) samples from A. glabripennis that fed on pin oak (Quercus palustris) (syringyl and guaiacyl lignin) and frass samples from Z. angusticollis that fed on ponderosa pine (Pinus ponderosa) (guaiacyl lignin only) were analyzed by TMAH...
thermochemolysis/GC/MS (Fig. 2A and B). Both digested samples exhibited significant lignin side-chain oxidation (two-tailed Student’s t test). For A. glabripennis-digested pin oak (n = 6), relative to undigested samples (n = 6), the G6/G4 ratio increased from 0.49 to 2.4 (P < 0.001) and the S6/S4 ratio increased from 0.26 to 2.4 (P < 0.0001) (Fig. 2A, inset). Although all degraded samples showed an increase in the G4/G6 ratio, the ratio varied depending on the preparation of the wood. Because degradation of wood occurs on the outer surfaces, methods resulting in greater sampling of the inaccessible interior (fine grinding of sample) resulted in greater G4 abundance (lowered G6/G4 ratios.) For Z. angusticollis-digested ponderosa pine (n = 5), the G6/G4 ratio increased from 0.58 to 2.0 (P < 0.001) (Fig. 2B, inset). Interestingly, this oxidation and depolymerization of lignin occurred at similar levels to but much more rapidly (hours vs. weeks) than that observed with white-rot fungal degradation of loblolly pine (16). Thus, this indicates that the extent of lignin modification by these two insect systems is equivalent to that observed with wood-degrading fungi.

Ring hydroxylation of guaiacyl units results in formation of syringyl units in lignin; an increase in syringyl compounds is indicative of ring hydroxylation. Hydroxylation can occur on intact lignin or side-chain–oxidized lignin. The diastereomeric pair of the enantiomers 1-(3,4-dimethoxy)-1,2,3-trimethoxypropane (S14/15) from S lignin or 1-(3,4,5-trimethoxyphenyl)-1,2,3-trimethoxypropane (G14/15) and S4/14/15 that result from this analysis. Asterisks represent 13C-labeled carbons resulting from 13C-TMAH analysis.

To determine whether demethylation of ring methoxyl groups had occurred and also to confirm that syringyl units found in the digested pine were derived from ring hydroxylation, we used 13C-TMAH thermochemolysis. Unlabeled TMAH cannot be used to assay for demethylation because it methylates all ring hydroxyls and they become indistinguishable from the original methoxyl groups (Fig. 1, reaction 3). With 13C-TMAH, the molecular weight of the demethylated product exhibits an increase in one mass unit for each site of demethylation (Fig. 1, reaction 3). Freshly collected A. glabripennis frass showed significant amounts of demethylation (n = 4) and compounds G4, S4, and G6 were isotopically enriched, whereas degraded wood (n = 4) showed no enrichment (Table 1). Inspection of 1-(3,4-dimethoxyphenyl)-1,2,3-trimethoxypropane (G14/15) and S14/15, which are proxies for the intact lignin polymer, showed little or no demethylation, consistent with demethylation occurring after side-chain oxidation. In Z. angusticollis frass samples containing syringyl lignin-to-guaiacyl lignin in the degraded wood (S/G = 2.53) compared with control wood (S/G = 2.07; P = 0.011, two-tailed Student’s t test). This decrease is not necessarily attributable to a lack of ring hydroxylation but, instead, could be attributable to preferential degradation of syringyl lignin over guaiacyl lignin. Syringyl lignin is known to be more easily degraded and depolymerized; thus, preferential removal of syringyl compounds would be expected, making hydroxylation difficult to detect (18).

Fig. 1. Lignin biodegradation reactions. Reactions include propyl side-chain oxidation/cleavage (reaction 1), ring hydroxylation (reaction 2), and demethylation (reaction 3). Reaction 1 is characteristic of white-rot fungus, whereas reactions 2 and 3 are characteristic of brown-rot fungus. Application of TMAH thermochemolysis to each of the degradative reactions is pictured following each reaction, showing chemical structures of compounds G4, G6, G14/15, and S14/15 that result from this analysis. Asterisks represent 13C-labeled carbons resulting from 13C-TMAH analysis.

Fig. 2. Side-chain oxidation and hydroxylation of wood after passage through insect guts. Chromatograms show increases in G6 and S6 in wood fed on by A. glabripennis (A) and Z. angusticollis (B) after passage through gut (red line = degraded [offset 5%]), black line = undegraded). Inset graphs show significant changes in ratio, or side-chain oxidation of guaiacyl (solid) and syringyl (hatched) lignin between degraded (U) and degraded (D) wood. Bars represent means ± 1 standard error. (B) In the Z. angusticollis chromatogram, hydroxylation of guaiacyl lignin can be seen through the production of S4 and S6 (from G4 and G6) in degraded samples, which were not present in the undegraded samples.
analysis provided no evidence of demethylation of G4. For G6, 13C-TMAH analysis actually showed a decrease in the relative percentage of demethylation (Table 1). Note that in guaiacyl lignin, there is only one site for ring demethylation, whereas in syringyl lignin, there is an additional aromatic methoxyl group; thus, syringyl-derived structures have one or two sites for demethylation. Data shown are the mean ± standard error. NP, not present in sample.

\( \frac{\text{Product No. OH}}{\text{(% yield on sample)}} \)

<table>
<thead>
<tr>
<th>Product</th>
<th>No. OH</th>
<th>Undegraded (%) yield on sample</th>
<th>Degraded (%) yield on sample</th>
</tr>
</thead>
<tbody>
<tr>
<td>G4</td>
<td>1OH</td>
<td>4.57 ± 1.42</td>
<td>8.28 ± 1.02*</td>
</tr>
<tr>
<td>G6</td>
<td>1OH</td>
<td>25.1 ± 2.1</td>
<td>35.6 ± 3.1*</td>
</tr>
<tr>
<td>G14</td>
<td>1OH</td>
<td>0.57 ± 0.41</td>
<td>1.34 ± 0.52</td>
</tr>
<tr>
<td>G15</td>
<td>1OH</td>
<td>1.21 ± 0.54</td>
<td>1.06 ± 0.51</td>
</tr>
<tr>
<td>S4</td>
<td>2OH</td>
<td>2.57 ± 0.19</td>
<td>2.56 ± 0.20</td>
</tr>
<tr>
<td>S4</td>
<td>2OH</td>
<td>0.52 ± 0.02</td>
<td>0.19 ± 0.09*</td>
</tr>
<tr>
<td>S16</td>
<td>2OH</td>
<td>6.70 ± 0.40</td>
<td>6.42 ± 0.07</td>
</tr>
<tr>
<td>S16</td>
<td>2OH</td>
<td>56.9 ± 2.2</td>
<td>70.2 ± 9.2</td>
</tr>
<tr>
<td>G14</td>
<td>1OH</td>
<td>1.42 ± 8.28</td>
<td>8.28 ± 3.77</td>
</tr>
<tr>
<td>G15</td>
<td>1OH</td>
<td>0.20 NP</td>
<td>29.4</td>
</tr>
<tr>
<td>S4</td>
<td>2OH</td>
<td>0.45 NP</td>
<td>3.38</td>
</tr>
<tr>
<td>S16</td>
<td>2OH</td>
<td>0.40 NP</td>
<td>25.6</td>
</tr>
<tr>
<td>S16</td>
<td>2OH</td>
<td>0.07 NP</td>
<td>5.35</td>
</tr>
</tbody>
</table>

Percentage of compounds with phenolic hydroxyl (1- or 2-OH) groups resulting from demethylation of ring methoxyl groups. In guaiacyl lignin, there is only one site for ring demethylation, whereas in syringyl lignin, there is an additional aromatic methoxyl group; thus, syringyl-derived structures have one or two sites for demethylation. Data shown are the mean ± standard error. NP, not present in sample.

\( ^*P < 0.05, \) significant changes between degraded and undegraded samples (two-tailed Student's \( t \) test).

\( ^{1}P < 0.001. \)

\( ^{2}P < 0.01. \)

\( ^{3}\)Nearly significant, \( P = 0.0508. \)

\( (n = 4) \), compared with undegraded wood \( (n = 4) \), 13C-TMAH analysis provided no evidence of demethylation of G4. For G6, 13C-TMAH analysis actually showed a decrease in the relative amount of the methylated versus demethylated product (Table 1). This is consistent with consumption of any original demethylated units (15). In contrast, the guaiacyl units from intact lignin (G14/15) of degraded wood were more extensively demethylated than fresh wood (Table 1), indicating that demethylation reactions occurred on the intact lignin polymer. The syringyl units generated from hydroxylation of guaiacyl units exhibited a high percentage of demethylation (Table 1). Note that in guaiacyl lignin, there is only one site for ring demethylation, whereas in syringyl lignin, there is an additional aromatic methoxyl group; thus, syringyl-derived structures have one or two sites for demethylation.

Although we have shown that lignin is degraded by both insect gut systems, the source of the enzymes/reagents responsible for these reactions is not known. Because fungi are the predominant degraders of lignin, we also investigated the potential role of fungi in these processes. In A. glabripennis, we analyzed total gut DNA of A. glabripennis larvae fed in Q. palustris for the presence of fungi using PCR amplification of the internal transcribed spacer (ITS) region with fungus-specific primers. Cloning and sequencing of 48 clones yielded only one strain of fungus, with all clones having an identical DNA sequence. Basic local alignment search tool analysis identified this sequence as an ascomycete fungus, falling into the same Fusarium solani/Nectria haematococca species complex as the Saperda vestita-related fungus. We also identified this fungal isolate from guts of A. glabripennis larvae infesting different tree species in New York and China (data not shown). F. solani/N. haematococca is a species complex known to have lignin-degrading capabilities (19), and although not closely related to basidiomycete wood-degrading fungi (Fig. 3), ascomycetes are known to degrade lignin and are responsible for wood decay referred to as “soft rot,” a process not well understood (20). Recent work by Shary et al. (20) using model lignin compounds demonstrated the ability of ascomycetes to catalyze propyl side-chain oxidation, similar to reactions shown in the A. glabripennis and Z. angusticollis systems. To substantiate participation of fungi in the A. glabripennis gut digestive system further, we performed metatranscriptomic analysis of the A. glabripennis gut microbial community. Both eukaryotic and prokaryotic transcripts were obtained, including fungal transcripts. These data are consistent with these fungi being metabolically active and playing a role in lignocellulose degradation in the beetle gut community.

Fig. 3. Soft-rot fungal species associated with the gut of A. glabripennis. Fungal community analysis of A. glabripennis gut contents yielded only a single strain of fungus based on ITS sequence, falling into the same species complex as a previously identified cerambycid related fungus. Phylogenetic analysis using the ITS region and performing a neighbor-joining analysis demonstrates that these fungi fall into the Ascomycota along with other known “soft-rot” degrading fungi (37), with both brown-rot and white-rot fungi being phylogenetically distinct. GenBank accession number for each sequence is listed next to the species name on the tree.
Discussion
In this study, we have shown definitive evidence for lignin degradation in the guts of two unrelated xylophagous species, *A. glabripennis* and *Z. angusticollis*. This work demonstrates lignin degradation in wood-feeding beetles; previous evidence of lignin degradation by termites was summarized by Breznak and Brune (7) as “scarce . . . and somewhat ambiguous.” These authors state that the ambiguity comes from the nature of the lignin preparation used for these studies and the “various conventions when referring to a substance as lignin.” Much of the “lignin” used in prior studies came from infusion of 1C-lignin precursors into growing plants with the assumption that the majority of the label is incorporated into polymeric lignin. The scepticism that termites degrade lignin was reinforced by the recent metagename sequencing of the higher termite species *Nasutitermes corniger* hindgut community. No genes encoding known lignin-degrading enzymes were found (21). However, the section of the hindgut investigated, although microbe-rich, is also anaerobic, excluding possible aerobic microbes that may be able to degrade lignin. Also, previous investigation of lignin modification by the lower termite *Cryptotermes brevis* noted only minor changes in the lignin molecule, with no evidence for modifications to side chains (22). In contrast to our work here, where the entire sample was analyzed, these earlier experiments characterized only the isolated (Björkman) lignin fraction from insect feces, limiting detection of monomeric units formed from side-chain oxidation.

Our results using TMAH thermochemolysis show that two different insect gut systems cause significant alteration of the lignin polymer. This analysis detects chemical modification of the entire lignin fraction of the sample; thus, changes detected can be attributed to the entire lignocellulose sample. All reactions known to occur in lignin biodegradation—side-chain oxidation, ring hydroxylation, and demethylation—occur in these insects. Side-chain oxidation, a reaction that leads to lignin degradation, is the predominant reaction observed with *A. glabripennis*, a process that can be catalyzed by white-rot fungi (3) and by ascomycetes (20). Many other species of cerambycids feed on decaying wood that is infested with wood-rot fungi. These “lower” cerambycids are not able to feed on fungus-free wood (23). Although *A. glabripennis* does not feed on wood infested with fungi, our data and those of others suggest that this insect has evolved a fungal symbiosis in the gut to allow digestion of living woody tissue. Interestingly, a fungal ascomycete, *Nectria haematococca*, was found in the gut of a closely related cerambycid species (Fig. 3), but the role of this fungus was not investigated (24).

For *Z. angusticollis*, our data show side-chain oxidation along with demethylation. Most notably, we detected ring hydroxylation as evidenced by formation of syringyl products from a wood that only contains guaiacyl lignin. All three reactions observed in the *Z. angusticollis* gut are more consistent with brown-rot fungal decay, but neither brown-, white-, nor soft-rot fungi has been shown previously to be associated with termite guts. The participation of fungi in *Z. angusticollis* wood degradation is a topic of ongoing research in our laboratories. Aside from fungi, the presence of aromatic degrading bacteria (actinomycetes) in termite guts has been well documented (25). Although these bacteria could possibly play a role in circumventing the lignin barrier (25, 26), the biochemical abilities of these bacteria are not as well defined as fungal systems. Researchers have yet to demonstrate extensive depolymerization of lignin by bacteria, and we suggest that bacteria are unlikely to be involved in side-chain oxidation. The presence of actinomycetes related to those harbored by termites was reported previously in the gut of *A. glabripennis*, and these bacteria may have similar aromatic degrading abilities (27).

Although fungal degradation of woody biomass has been viewed as the paradigm for lignocellulose degradation, our work here shows that lignin degradation takes place in an ecosystem in which it was previously thought not to occur. These two insect systems displayed similarities to the pathways described for white-, brown-, and soft-rot fungi, but they are instead occurring in insect-microbial systems rather than by the action of a single fungal species. Known lignin-degrading pathways are aerobic processes. Despite the belief that termites ferment cellulose anaerobically, aerobic regions of termite guts, in addition to those of other insects, are known to exist in which large populations of aerobic microbes are housed, which can outnumber anaerobic microbial populations in the gut in some cases (28–30). Although we cannot directly measure a rate of lignin degradation, or identify specific genes or organisms responsible for this degradation at this time, the striking modifications to lignin detected in wood after passage through these insect guts is quite astonishing in regard to the short period that wood is held in the insect digestive system (only several hours), compared with the weeks for which similar degradation has been reported for wood-rot fungal systems (15, 16). Being controlled systems, it is likely that these insect guts have evolved optimal conditions for lignocellulose degradation that cannot be obtained in free-living systems. Discovery that these gut microbial communities can degrade lignin provides considerable enthusiasm for bio-prospecting these systems, through metagenomic or proteomic approaches, for new lignin-degrading enzymes and systems. Further understanding of these systems will contribute to our understanding of the ecological roles that these insects play in carbon cycling in natural systems.

Materials and Methods

Sample Collection and Preparation. *A. glabripennis* and *Z. angusticollis* were obtained from colonies maintained at the Department of Entomology, Pennsylvania State University. *A. glabripennis* adults were allowed to oviposit into four potted pine oak (*Q. palustris*) trees in a quarantine greenhouse, and larvae were allowed to develop in trees for 6 weeks. At this point, trees were cut down, larvae were removed, and fresh frass samples were collected from each insect, which were immediately frozen in liquid nitrogen to halt wood degradation. Wood samples adjacent to but not in contact with larval galleries were also collected as control samples. In total, 10 frass samples and accompanying control wood blocks were collected. For termites, frass collection, five pairs of *Z. angusticollis* workers were removed from the colony and placed into cups containing 1 × 1 × 0.4 cm of ponderosa pine (*P. ponderosa*) blocks and allowed to feed for 4 days. During this time, frass was collected from each cup every 12 h and frozen. Samples were then pooled over the 4 days to provide enough material for analysis. Five control wood blocks were placed in additional cups, and a small sample was collected from them during each collection point and pooled. Frass samples and control wood tissue were lyophilized overnight. For unlabeled TMAH performed off-line, frass samples were not ground before analysis, but for online 13C-TMAH, samples were ground to a fine powder utilizing a freezer mill before being used for analysis. The percentage (%) of organic carbon content was determined for each sample using a Carlo Erba 1108 elemental analyzer (Carlo Erba) to normalize quantititative TMAH thermochemolysis analysis results.

Standard TMAH Thermochemolysis Analysis. Tetramethylammonium hydroxide–thermochemolysis experiments performed as modified by Frazier et al. (31). Approximately 0.5 mg of each sample and 150 µl of TMAH (25 wt % TMAH in methanol; Fisher Scientific) were placed in borosilicate glass tubes and mixed by vortexing for 30 sec. The samples were dried under nitrogen and vacuum-sealed on a manifold before being baked at 250°C for 30 min. The sealed tubes were cooled before cutting the glass and adding an internal standard, linolenic acid methyl ester (Sigma Aldrich) at 54.2 ng/µl in ethyl acetate. Samples were extracted from tubes by washing the sides of the tubes with ethyl acetate (Fisher Scientific). All washings (approximately 1 ml per sample) were filtered through glass wool, combined, and concentrated to 200 µl under a stream of nitrogen. The samples were analyzed by capillary GC-MS on a Hewlett Packard 6890 series GC system connected to a Pegasus III time-of-flight mass spectrometer (LECO Corporation). The column used for the GC
separation was a 15-m × 0.25-mm i.d. fused silica capillary column with a film thickness of 0.25 μm (5% phenylpolysiloxane – 95% methylpolysiloxane, Restek RtX-S). The samples (1 μl) were injected onto a split/splitless injector operating in splitless mode at a temperature of 280°C. The column, using helium as the carrier gas, was run with a constant flow (1.0 ml/min), and the oven temperature was held initially at 50°C for 2 min, then ramped from 50°C to 300°C at 15°C per minute, and then held at 300°C for 6.33 min, giving a total run time of 25 min. The injection mode on the mass spectrometer was electron ionization at 70 eV, and the ion source temperature was 200°C, whereas the transfer line temperature was 280°C. Mass spectra were collected at a rate of 20 spectra per second after the 120-sec solvent delay. Masses were acquired between m/z 45 and 500, and data acquisition and analysis were performed by use of the LECO Pegasus III software (version 3.22). Chromatograms obtained from GCTOF-MS were integrated using masses 166, 196, and 226, which are specific for G4, G6, S4, and S6. This was done to remove small peaks that coelute with G6 from the total ion count. After integration, the peak area of each of these four compounds was determined, and G6-G4 and S6-S4 were calculated (32).

14C-TMAH Thermochemolysis Analysis. 13C-labeled TMAH was synthesized as previously described (33). 13C-TMAH thermochemolysis was performed in-line using a Shimadzu Pyr-4a pyrolysis unit, which also doubles as a stand-alone injection port (modified from Filley et al. [34]). For each treatment, three replicate samples were analyzed from unique insect pools along with three corresponding controls. A total of 150 μg of each sample was weighed into a 3 × 3-mm platinum bucket that contained eicosane internal standard. Next, 3 μl of a 25-μl % solution of 13C-TMAH in water was added to the bucket and placed in the sample holder on the top of the Pyr-4a pyrolysis unit. The TMAH was allowed to soak into the sample for 5 min, after which the bucket was dropped into the pyrolysis unit, which was isothermal at 350°C. The injector base of the Py-4a was maintained at 320°C with a split flow of 25/1. Chromatography and mass spectral analysis were performed on a Shimadzu GC17A interfaced to a Shimadzu QP5050A quadrupole mass spectrometer collecting in mass range 40–550 amu. Chromatographic separation was performed on a fused silica column (Restek RTX-5 [Restek, 30 m, 0.25 mm i.d., 0.25-μm film thickness]). The oven was temperature programmed from 60°C (1 min stationary) to 300°C at 7°C per minute and maintained at that temperature for 15 min. Chromatograms obtained from this analysis were analyzed [as described previously (16, 34)] for S/G to measure hydroxyl functionality and % 13C-labeled methyl groups on G4, G6, G14, G15, S4, and S6 to determine the extent of demethylation of each of these compounds in each sample.

Fungal Community Analysis of the *A. glabripennis* Gut. Adult *A. glabripennis* were allowed to oviposit onto *Q. palustris* trees, and larvae were allowed to grow and develop for 90 days as described previously. After 90 days, each tree was dissected and living larvae were collected for gut community analysis. Ten living larvae were used for gut dissections, which were performed in a laminar flow hood to maintain sterility using sterile dissection tools. Larvae removed from trees were separately and dissected and contained within 1 of 6 removal from 6 trees. Larvae were surface sterilized in 70% ethanol for 1 min and rinsed in sterile water before dissection. Whole guts were dissected by cutting the cuticle open laterally, ligating the gut at the anterior midgut and posterior hindgut, and carefully transferring the entire gut into a sterile microcentrifuge tube. Ten guts were pooled into a single microcentrifuge tube for DNA extraction for each tree species repeated each strain variation within trees. Total DNA was extracted using the FastDNA SPIN for Soil Kit (MP Biomedicals) using the FastPrep Instrument for tissue homogenization following the manufacturer’s protocol. This kit was used because of the complexity of the *A. glabripennis* gut contents (containing wood, bacteria, and fungi) to ensure complete DNA extraction from all organisms. A control DNA extraction was also performed by using the sterile water rinse to confirm that no contaminating DNA was extracted. DNA concentration was determined by absorbance at 260 nm and stored at −20°C until use. The fungal ITS region was amplified using PCR from total gut DNA extractions. Fungal ITS primers IT51 (5'-TCC GTA GGT GAA CCT GCG G-3') and IT54 (5'-TCC TTC OGT TAT GAT TAC GGC-3') were used to amplify 500 bases of the ITS region. Pyrosequencing reactions were performed in 25-μl volumes with the following components: 5 μl of 5× GoTaq green reaction buffer, 0.5 μl of GoTaq DNA polymerase (1.25 U, Promega), 2 μl of 10 μM forward primer (IT51), 2 μl of 10 μM reverse primer (IT54), and 20 ng of template DNA. Polymerase chain reaction conditions were 95°C denaturation for 3 min, 25 cycles of 95°C for 30 sec, 55°C for 1 min, and 72°C for 1.5 min, with a final extension at 72°C for 5 min. Control DNA extraction samples also underwent PCR to ensure that there was no non-gut contaminating DNA during extraction. Polymerase chain reaction amplification revealed a single band from ITS amplification, but that single band may contain more than one fungal sequence. To resolve this, the PCR product was ligated into pCR 2.1 TOPO vector (Invitrogen) following the manufacturer’s protocol. The vector was then transformed into chemically competent Escherichia coli cells (TOP10, Invitrogen Corp.) by heat-shock, and clone libraries were created. Insert DNA from clones was amplified from the M13 priming site. Each insert was sequenced using direct M13 priming. Five for each clone, followed with the following components for each reaction: 5 μl of 5× GoTaq green reaction buffer, 0.5 μl of GoTaq DNA polymerase (1.25 U, Promega), 2 μl of 10 μM forward primer (M13 Universal), and 2 μl of 10 μM reverse primer (M13Rev). Individual colonies were picked from the clone library using a sterile pipette tip and immersed into the PCR mix to allow the bacteria cells to be added to the PCR. The PCR program had an initial 95°C denaturation for 10 min to rupture bacteria cells, followed by 30 cycles of 95°C for 30 sec, 55°C for 1 min, and 72°C for 1.5 min, with a final extension at 72°C for 5 min. Four microliters of the PCR product was cleaned up for sequencing by addition of 0.8 μl of ExoSAP-IT (USB Corporation) and incubating the sample at 37°C for 15 min, followed by 80°C for 30 min. Two microliters of this reaction was then used to sequence from the forward direction, from the M13 universal priming site, and 2 μl was used for the reverse from the M13 reverse site. Sequencing using the BigDye Terminator (Applied Biosystems) method was performed at the Penn State Nucleic Acid Facility. Internal transcribed spacer sequences were constructed over the regions amplified based on consensus of the forward and reverse sequences for each clone analyzed. Alignment of forward and reverse sequences was performed using MEGA 4.0.1 (36), and the vector sequence was removed and a consensus sequence was created. All sequences from the clone library were identical; thus, a representative sequence (deposited into GenBank, accession no. EU882818) was used for all the downstream analyses. The vector was then transformed into chemically competent *E. coli* cells (TOP10, Invitrogen Corp.) by heat-shock, and clone libraries were created. Insert DNA from clones was amplified from the M13 priming site. Sequences were aligned by ClustalX (Conway Institute, UCD Dublin) in MEGA, and from this alignment, a neighbor-joining tree was created and clustered with a bootstrap test (50 replicates) (38, 39).

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